

Combined effect of *Bacillus cereus* CPOU13 and *B. subtilis* SPC14 on polycyclic aromatic hydrocarbons degradation *in vitro*

Poornachander Rao M.1*, Anitha Y.2, Satyaprasad K.3

¹Department of Botany, Government Degree College, Mahabubabad, Warangal (Dist.), Telangana State. ²CRIDA (ICAR, New Delhi), Hyderabad, Telangana State. ³Professor in Botany (Rtd.), Department of Botany, Osmania University, Hyderabad

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Abstract: Biodegradation of polycyclic aromatic hydrocarbons (PAHs) with suitable bacterial strains conveys much interest in recent years. We studied biodegradation of PAHs namely phenanthrene, anthracene and pyrene using two efficient PAHs degrading strains, *B. cereus* CPOU13 and *B. subtilis* SPC14 *in vitro* experiments. Standard HPLC chromatograms for phenanthrene, anthracene and pyrene was geprately based on HPLC peak area values and Retension time of known concentrations of the test PAHs and using software, 'Origin 6.0'. Biodegradation of PAHs mixture containing phenanthrene, anthracene and pyrene was studied *in* for 13 days. We found that the combination of bacterial strains, *B. cereus* CPOU13 and *B. subtilis* SPC14 degraded high amounts of phenanthrene, anthracene and pyrene in 13 days of incubation. The recorded degradation percentages of phenanthrene, anthracene and pyrene were 85.31, 92.82 and 85.70 respectively. Concentration of phenanthrene was degraded from 217µg/5ml to 31.9µg/5ml. Concentration of pyrene was degraded from 233µg/5ml to 33µg/5ml in 13 days of incubation. We also observed biodegradation of phenanthrene, anthracene and pyrene was degraded from 213µg/5ml to 33µg/5ml in 13 days.

Key words: Biodegradation; PAHs; B. cereus CPOU13; B. subtilis SPC14; phenanthrene; anthracene; pyrene

Introduction

Polycyclic aromatic hydrocarbons (PAHs) are ubiquitous environmental contaminants that formed by the incomplete combustion of organic materials, such as wood or fossil fuels. PAHs consist three or more benzene rings, at least two are fused with two neighboring rings and shares two adjacent carbon atoms (1). There are thousands of PAHs in the environment but in practice PAHs analysis is restricted to determination of 6 to 16 compounds. Individually all PAHs differ substantially in their physical and chemical properties (2). The ubiquitous occurrence of PAHs is largely due to their formation and release in all processes of incomplete combustion of organic materials and these organic contaminants are resistant to degradation, can remain in the environment for long periods (3, 4).

Presence of PAHs in the environment at high levels exerts negative impact on all life forms due to their carcinogenic, mutagenic, teratogenic and phytotoxic properties and this enforced the scientific world to concentrate on the immediate actions for remediation of PAHs (5-7). Biodegradation with suitable bacterial strains conveys much interest in recent years due to its low cost, effectiveness, restoration of medium it operated and possesses several where advantages. Few bacterial strains act as key participants in PAHs biodegradation as they possessing special enzymes. Strains suitable for PAHs biodegradation studies are frequently isolate from their native contaminated sites and this may be due to their long term exposure and tolerance

*Corresponding Author:

Dr. M. Poorna Chander Rao Asst. Prof. in Botany, Govt. Degree College, Mahabubabad, Warangal (Dist.), Telangana State. towards PAHs (8). However, consortia or multibacterial strains of PAHs degrading strains are more effective in biodegradation as compared to single strains (9, 10). In the present investigation, we used two effective PAHs degrading bacteria strains namely *B. cereus* CPOU13 (11) and *B. subtilis* SPC14 (12) those isolated from two different PAHs contaminated sites to study degradation of PAHs phenanthrene, anthracene and pyrene *in vitro* experiments.

Materials and Methods

Construction of standard chromatogram for HPLC studies:

Method for the construction of standard chromatograms for PAHs using high performance liquid chromatography (HPLC) was adopted from Boonchan (13). Phenanthrene, anthracene and pyrene with 99% purity purchased from Sigma-Aldrich, USA and used throughout the experimental work as test PAHs. Accurately weighed 100 mg of phenanthrene, anthracene and pyrene transferred to a 100 ml volumetric flask and 2-3 ml of acetonitrile was added to dissolve. Solution was diluted with acetonitrile to obtain 1000 mg/L stock solution. The stock solution further diluted to obtain the concentrations of 25, 50, 100, 200 and 250 ppm and the samples run in HPLC. Standard chromatograms for phenanthrene, anthracene and pyrene were plotted separately using software, 'Origin 6.0' by retrieving the peak area values at respective retention times of each PAH.

HPLC analysis:

High performance liquid chromatography (HPLC) studies were conducted with a reverse phase HPLC (SHIMADZU, model RF-10AXL). The instrument consists of dual pump system and connected with UV detector (SPD-20A). Instrument was equipped with column C 18 (250 mm x 4.6 mm, 5 A° particle size) of Phenomenex Co. Mobile phase was consisted of 75% acetonitrile and 25% of de-ionized water. Detector was set at 250 nm and mobile phase was maintained at flow rate of 0.8 ml/min in isocratic mode. 20 µl of sample was injected into HPLC with a HPLC injector (Rheodine injector) that prior filtered with 0.22 µm syringe filters. Data of each peak on HPLC chromatogram was analyzed using chromatography software 'LC Solutions'.

Degradation of polycyclic aromatic hydrocarbons (PAHs) *in vitro*:

Degradation of PAHs mixture containing phenanthrene, anthracene and pyrene was studied by the method of Moody (14). 250 ml of Minimal salt medium (MSM) was prepared in 1000 ml conical flasks and enriched with phenanthrene, anthracene and pyrene (each 50 ppm) and made 150 ppm. One ml of exponential growth phase culture (approximately 107 colony forming units) of strains (B. cereus CPOU13 and B. subtilis SPC14) were added separately to MSM flasks and in control flasks no culture was added. The cultures were incubated under standard growth conditions (at 24°C on a rotary shaker at 150 rpm speed) for 13 days in the dark. A 5 ml of aqueous portion was withdrawn from each flask at regular time intervals of 0, 1, 2, 3, 7, 8 and 13 days and extracted with three equal volumes of ethyl acetate after adjusting pH to 2.5 using 1N HCl and the step of extraction was repeated thrice. Each time a pinch of anhydrous disodium sulphate (Na₂SO₄) was added to remove residual water content from the samples. Extract was concentrated using rotary evaporator under reduced pressure at 34°C under vaccum conditions. Finally, the samples were dissolved in 3 ml of acetonitrile (ACN) and preserved in a refrigerator at 4°C for HPLC studies as described earlier.

Results

Preparation of standard chromatograms for PAHs:

Standard chromatograms for phenanthrene, anthracene and pyrene were plotted by running HPLC with the known concentrations 25, 50, 100, 200 and 250 ppm separately. Each PAH formed a peak at constant retention time and peak areas were increased with the increase in concentration. Based on the values of peak areas against known concentrations (25, 50, 100, 200 and 250 ppm) standard chromatograms for each PAH were plotted separately using the software 'Origin 6.0'.

Table 1: HPLC peak area values of phenanthrene, anthracene and pyrene at known concentrations

Sl. No.	Concentration	Peak Area Values					
	(ppm)	Phenanthrene	Anthracene	Pyrene			
1	25	25.880	20.688	10.233			
2	50	221.241	80.641	40.498			
3	100	625.788	260.992	120.054			
4	200	855.055	640.167	280.330			
5	250	1017.880	918.880	420.777			

Phenanthrene peak was appeared at 13 min retention time on HPLC chromatograms. Anthracene peak was recorded at retention time of 13.5 min. Pyrene was formed a peak at the retention time of 16.5 min. The standard chromatograms were constructed separately for each PAH to study PAHs concentrations in the standard solution based on their retention time and peak area (Fig 1, 2, 3). A linear standard graph obtained for each PAH and their correlation coefficient values are 0.9935 for phenanthrene, 0.9916 for anthracene and 0.9865 for pyrene. The value of the correlation coefficient (R²) obtained for each calibration chromatogram shows that the correlation between relative peak area and concentration was linear and reproducible within selected concentration range. The 'y'- value representing linear correlation equation for phenanthrene, anthracene and pyrene were 3.921x, 3.9528x and 1.7749x respectively. Thus, the data obtained from standard chromatogram was reliable and accurate (Fig 1, 2, 3).

Figure 1: Standard graph for the HPLC chromatograms of phenanthrene

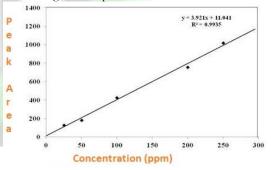


Figure 2: Standard graph for the HPLC chromatograms of anthracene

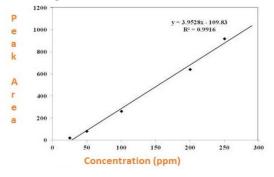
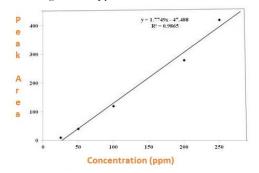


Fig. 3: Standard graph for the HPLC chromatograms of pyrene

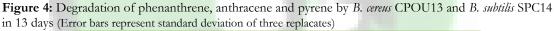


Degradation of PAHs by *B. cereus* CPOU13 and *B. subtilis* SPC14 *in vitro*:

Combined biodegradation capability of *B. cereus* CPOU13 and *B. subtilis* SPC14 was assayed *in vitro* conditions on MSM enriched with a mixture of phenanthrene, anthracene and pyrene each at 50 ppm concentration. This was conducted by high performance liquid chromatography (HPLC) with C 18 column. Degradation of PAHs recorded at regular time intervals of 0th, 1st, 2nd, 3rd, 7th, 8th and 13th days of incubation. The observed retention peak areas of phenanthrene, anthracene and pyrene were analyzed by software 'LC-solutions'. Based on retention times and respective peak area values, concentrations of PAHs in the samples were determined with the help of standard chromatograms plotted for known concentrations of PAHs. Further, based on the concentration of each PAH remained in respective samples degradation percentages were calculated.

The combination of bacterial strains, *B. cereus* CPOU13 and *B. subtilis* SPC14 resulted in high degradation performance for phenanthrene, anthracene and pyrene in 13 days of incubation (Fig. 4, 5). Degradation percentages and concentration depletions of phenanthrene, anthracene and pyrene by the strains showed a gradual increase from the 1st day to 13th day. The results are presented in Table 2.

Degradation of phenanthrene was recorded up to 85.31% in 13 days as its initial concentration degraded from $217\mu g/5ml$ to $31.9\mu g/5ml$. Degradation of anthracene was recorded up to 92.25% as its concentration declined from $211\mu g/5ml$ to $16\mu g/5ml$ in 13 days. Degradation of pyrene was reached to 85.70% as its initial concentration depleted from $233.14\mu g/5ml$ to $33\mu g/5ml$ in 13 days of incubation.



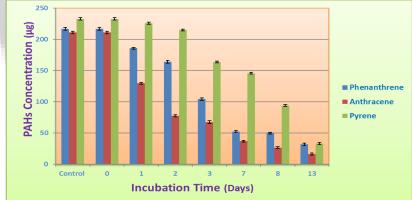


Table 2: Degradation of phenanthrene, anthracene and pyrene by combined treatment of *B. cereus* CPOU13 and *B. subtilis* SPC14

Day	Phenanthrene		D 1.	Anthracene		Desiredat	Pyrene		D 1.
	RT Peak	Conc. (µg/5ml)	0	RT Peak	Conc. (µg/5ml)	ion (%)	RT Peak	Conc. (µg/5ml)	—Degradat ion (%)
Control	880.750	217 ± 2.07	0	678.142	211±2.03	0	380.241	233 ± 2.08	0
0 th Day	880.750	217 ± 2.07	0	678.142	211 ± 2.03	0	380.241	233 ± 2.08	0
1 st Day	754.98	186±1.54	14.39	421.073	130 ± 1.54	38.63	372.91	226±1.63	3.00
2nd Day	665.122	164 ± 2.08	24.41	252.232	77 ± 1.56	63.29	353.068	215±1.39	7.65
3rd Day	416.597	104 ± 2.00	51.92	213.069	67 ± 2.07	68.07	272.885	164±1.58	29.64
7th Day	212.248	52±1.54	75.82	120.79	37±1.61	82.51	238.017	146 ± 1.64	37.42
8th Day	200.114	45±1.51	77.04	80.901	27 ± 1.48	87.39	153.226	94±1.39	59.60
13th Day	122.969	31.9±2.09	85.31	48.73	16±1.62	92.25	52.267	33±1.59	85.70
	Control 0 th Day 1 st Day 2 nd Day 3 rd Day 7 th Day 8 th Day	Day RT Peak Control 880.750 0 th Day 880.750 1 st Day 754.98 2 nd Day 665.122 3 rd Day 416.597 7 th Day 212.248 8 th Day 200.114	$\begin{array}{c c c c c c c c c c c c c c c c c c c $	$\begin{array}{c c c c c c c c c c c c c c c c c c c $	$\begin{array}{ c c c c c c c c c c c c c c c c c c c$	$\begin{array}{ c c c c c c c c c c c c c c c c c c c$	$\begin{array}{c c c c c c c c c c c c c c c c c c c $	$\begin{array}{c c c c c c c c c c c c c c c c c c c $	$\begin{array}{c c c c c c c c c c c c c c c c c c c $

(RT= Retention; ± represents standard deviation of three replicates)

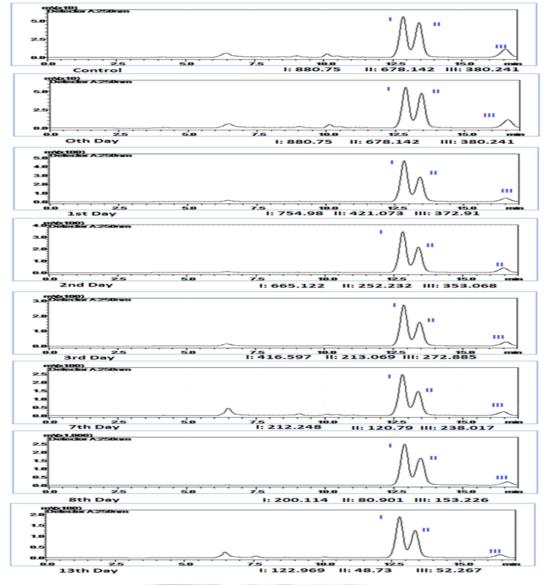


Figure 5: HPLC chromatograms showing the degradation of PAHs *in vitro* by *B. cereus* CPOU13 and *B. subtilis* SPC14

(I= Phenanthrene; II= Anthracene; III= Pyrene)

Discussion

Bacterial strains generally evolve to utilize more carbon molecules than one type of carbon molecules for their growth and metabolism. This flexibility for nutritional resources helps the bacteria to strive many nutritional constraint conditions. As several researchers suggested many bacterial strains utilize more than one type of PAHs as nutritional source in addition to commonly utilized carbohydrates. Bacterial strains acquire these qualities by secreting few specific enzymes that can catalyze many PAHs (15). These abilities of bacteria result in degradation or mineralization of PAHs on media components. However utilization of PAHs as nutritional sources in vitro experiments believed to be limited because of their hydrophobicity, crystalline and

other related properties which decrease their availability to degrading bacterial strains (16, 17).

Different analytical methods are adoptable to determine PAHs quantities in liquid samples during biodegradation. Reverse phase high performance liquid chromatography (RP-HPLC) is one of the prime methods used because it grants several advantages including sensitive and selective detectors, easy sample preparation, and less risk of solute degradation.

In our study phenanthrene, anthracene and pyrene formed different peaks in HPLC chromatograms at respective retention times (RT) which based on their respective number of rings and molecular weights. Sequential appearance of test PAHs on chromatogram was phenanthrene, anthracene and pyrene. These results are in agreement with the results of Haritash and Kaushik (18). Anthracene formed strong signals and high peaks in respective HPLC chromatograms due to their linear structure and high UV detection (19). The retention times observed for phenanthrene, anthracene and pyrene in HPLC chromatograms are 13, 13.5 and 16.5 minutes respectively. We also observed that peak areas of test PAHs increased with their concentration. Plotted standard chromatograms for each PAH had above the 0.9 correlation coefficient value. Hence, these results are reliable and amenable for determination of unknown concentrations of phenanthrene, anthracene and pyrene.

Strains suitable for biodegradation studies are generally isolate from their native contaminated sites. In more extent, consortia of PAHs degrading bacteria are more effective in degradation as compared to selected single strains (20). Yu (21) investigated biodegradation of mixture PAHs consisting of fluorene, phenanthrene and pyrene using bacterial consortium made up of three strains of *Rhodococcus* sp., *Acinetobacter* sp. and *Pseudomonas* sp. and reported that addition of the bacterial consortium significantly enhanced the efficiency of fluorene and phenanthrene biodegradation. In the present study we used two potent PAHs degrading bacterial strains that isolated from two PAHs contaminated sites.

Nasseri (22) reported that consortium of two bacterial isolates of Pseudomonas species showed maximum biodegradation of phenanthrene than the individual treatments. Leblond (23) also reported consortium of four different bacterial strains was more effective than single strain in PAHs biodegradation. Churchill (24) suggested that use of single bacteria for biodegradation took much more time than the time taken by a mixture of culture. This effect may be due to the ability of certain strains in removal of intermediates produced by other members of consortium that facilitates additional PAHs removal (25). In the present investigation, degradation of PAHs mixture (phenanthrene, anthracene and pyrene) by B. cereus CPOU13 and B. subtilis SPC14 was studied in vitro for 13 days. The combined treatment recorded high degradation rates of phenanthrene, anthracene and pyrene. Degradation percentages of phenanthrene, anthracene and pyrene in 13 days of incubation were 85.31, 92.82 and 85.70 respectively. These results indicating that combination of these strains is compatible and results in high degradation percentages.

Guo (26) suggested that certain defined bacterial consortia are very effective in biodegradation than the other consortia partners and isolated strains. They studied degradation of multi-PAHs by bacterial consortium and reported that among three bacterial strains, SPNT (Paracoccus versutus) degraded phenanthrene, fluoranthene and pyrene completely within 10 days. Strains, HCSS (Rhodococcus opacus) and MWFG (Sphingomona sp.) recorded lower degradation percentages for all three PAHs, especially pyrene, which had only 20% degradation. Similarly, in the present study combination of two PAHs degrading strains recorded high degradation of phenanthrene and anthracene in 13 days of incubation. Pyrene, a high molecular weight PAH also degraded to high percentage. The present results indicating that efficacy of PAHs degradation is high and can be improved by adding more bacterial strains. More efforts have to pay for the construction of highly PAHs degrading consortium for 100 percent PAHs degradation in short periods.

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